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# Distribution of soil nutrients, extracellular enzyme activities and microbial communities across particle-size fractions in a long-term fertilizer experiment

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#### ABSTRACT

Soils were particle-size fractionated to evaluate changes in carbon and nitrogen contents, enzyme activities and microbial community composition in response to 33 years of fertilization. This study focused on yellow-brown paddy soil and the particle-size fractions of >2000, 2000-200, 200-63, 63-2 and 2-0.1 µm. Microplate fluorometric assays and phospholipid fatty acid analysis (PLFA) were used to determine soil biological characteristics under no fertilizer (control, CK), fertilizer N(N), fertilizer N and P (NP), fertilizer N, P and K (NPK), organic manure plus fertilizer N, P and K (NPKM) and organic manure (M) treatments. The results showed that fertilizer and soil fraction individually and interactively (P < 0.05) affected soil C, N contents, enzyme activities and microbial communities except for  $\alpha$ -glucosidase activity, bacterial relative abundance and the G+:G- ratio. Particularly, organic treatments significantly increased soil organic carbon (SOC) and total nitrogen (N) contents of all five fractions. The highest C and N contents and enzyme activities were observed in the 200-63 µm fraction, except for phosphatase and sulfatase, which showed the highest activities in the 2–0.1 µm fraction. The highest activities of  $\beta$ -glucosidase,  $\beta$ -cellobiosidase,  $\alpha$ -glucosidase, aminopeptidase, phenol oxidase and peroxidase in each fraction were obtained in the organic treatments (NPKM and M). Activities of phosphatase, sulfatase, N-acetyl-glucosaminidase and  $\beta$ -xylosidase in the 2000–200  $\mu$ m fraction were highest under NPK treatment. PLFA analysis showed that the >63 µm fraction contained higher abundance of total PLFAs than that in the 63-0.1 µm fraction. Organic treatments significantly enhanced total PLFAs abundance in >2000 μm fraction, but decreased PLFAs abundance in the 2000-200 μm fraction compared with the NPK treatment. Larger fractions (>2000 µm and 2000-200 µm) held relatively lower G+:G- ratios and higher fungi:bacteria ratios, which indicated better soil conditions in these fractions. Principal component analysis showed a smaller variability of microbial community composition among treatments than particle-size fractions. Most treatments of larger fractions (>2000 µm and 2000-2 µm) were well separated from the other fractions. Redundancy analysis showed total N, C:N ratio, phosphatase, sulfatase, N-acetyl-glucosaminidase and β-cellobiosidase activities significantly affected the composition of the microbial community. Significant correlations were also obtained between enzyme activities with SOC, total N and C:N ratio. We concluded that the long-term application of organic fertilizers contributed to improvements in the soil organic carbon and total nitrogen and most of the enzyme activities, especially for the 200-63 µm fraction, along with abundant and diverse microbial community composition in larger particles.

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# 1. Introduction

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http://dx.doi.org/10.1016/j.apsoil.2015.05.005 0929-1393/© 2015 Elsevier B.V. All rights reserved. Soils consist of particles of sand, silt and clay which bound into aggregates of various sizes by organic and inorganic agents. The distribution and stability of aggregates, and of the pores within and between them, affect soil properties and the composition and activity of soil biotic communities (Tisdall, 1994; Mikha and Rice, 2004). At the same time, aggregate formation and stabilization are







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affected by various factors, including climate conditions, mineral composition and types and amount of soil organic carbon (SOC) (Degens and Harris, 1997; Denef and Six, 2005). The main agents of aggregate formation and stabilization are organic materials, including persistent cementing agents, such as humic matter involved in stabilizing microaggregates, and transient bonding agents (e.g., polysaccharides derived from plants and micro-organisms) as well as temporary binding agents (e.g., fungal hyphae, fine roots, bacterial cells) related to formation and stabilization of macroaggregates (Tang et al., 2011). Therefore, the determination of SOC and the soil nitrogen content (total N) within different particle-size fractions is important for evaluating the effects of different fertilizer managements on soil quality.

Continuous use of imbalanced fertilizers under intensive ricewheat cultivation in the Yangtze Plain has adverse impacts on the soil. Integrated nutrient management practice is seen as a viable option in restoring the soil physical structure and chemical fertility (Gattinger et al., 2012), improving soil organic C and therefore, sustaining the system productivity (Seufert et al., 2012) and reducing global warming potential (Cavigelli et al., 2013). The scientific literature is replete with studies on the effects of organic matter on soil structure and other properties, but frequently at the bulk soil scale. Little attention has been given to understanding the soil mechanism properties at the particle-size scale, which might manifest quite differently, but largely influence the bulk soil properties (Das et al., 2014). Therefore, the specific locations of enzyme activities within the soil matrix have attracted attention, especially as the area of these enzyme activities is affected by SOM quality (Kandeler et al., 1999) and turnover (Stemmer et al., 1998). Additionally, extracellular enzyme activities in soils play an important role in the degradation of polymeric material and the supply of low molecular weight substrates to microorganisms (Nannipieri et al., 2012). To date, it is unknown how soil nutrients and extracellular enzyme activities are distributed across particlesize fractions after different fertilizer treatments in a long-term experiment.

Definitely, balanced fertilizer application can influence both soil organic matter content and quality, substantially increasing microbial diversity in soils, thus improving soil functions and increasing the resistance of soil to environmental stress (Wardle and Ghani, 1995; Pankhurst et al., 1996; Kennedy and Gewin, 1997; Degens et al., 2000; Nsabimana et al., 2004). Microbial communities in soils produce extracellular enzymes to acquire energy and resources from complex biomolecules in the soil environment (Burns, 1982). These enzymes are of interest on an ecosystem scale because they catalyze important transformations in the carbon (C), nitrogen (N) and phosphorus (P) cycles (Wallenstein and Burns, 2011). It is well known that the addition of easily decomposable substrates to soil rapidly stimulates the soil microflora, resulting in a significant increase in aggregate stability (De Gryze et al., 2005; Abiven et al., 2007). Denef et al. (2001) and De Gryze et al. (2005) showed that fungi significantly affected macroaggregates formation. Besides the physical effects of enmeshment of macroaggregates by hyphae, extracellular polysaccharides can be produced by hyphae, attaching microaggregates and binding them together into stable macroaggregates (Neufeldt et al., 1999). Additionally, the hydrophobicity of microbial extracellular polysaccharides contributes to the stabilization of macroaggregates by decreasing their wettability (Liu et al., 2010). Besides fungi, bacteria also exude extracellular polysaccharides to bind soil particles and increase inter-particle cohesion (Degens and Harris, 1997). Increases in the fungi: bacteria ratio have been linked to increases in soil C and the C:N ratio across landscapes (Fierer et al., 2009 De Vries et al., 2012) and in response to organic amendments (Bernard et al., 2012). Other studies have shown increases in phospholipid fatty acid biomarkers for arbuscular mycorrhizal fungi (AMF) in response to long-term organic management (Bossio et al., 1998; Moeskops et al., 2010, 2012). Studies also supports that fungal communities has been suggested as a means of increasing agroecosystem N retention and other functions (De Vries and Bardgett, 2012; Jackson et al., 2012). However, the relationships among soil C and N pools, extracellular enzyme activities and microbial community composition and structure in soils amended with inorganic and organic fertilizers are still poorly understood.

As stated previously, aggregate formation could be influenced by many different factors, so it is still necessary and meaningful to conduct experiments in yellow–brown paddy soil, which is a typical paddy soil in the Yangtze Plain of China. Our objective was to elucidate the distribution of organic C and total N, enzyme activities and microbial community composition within different particle-size fractions to determine the significant positive effects of adequate fertilization. It is thus reasonable to predict that the application of balanced fertilization in combination with organic amendments to a long-term rice–wheat rotation would significantly impact aggregation, increase extracellular enzyme activities and enrich the microbial community, thereby resulting in a better soil structure and a higher soil quality for crops.

# 2. Materials and methods

# 2.1. Field design and sampling

The long-term field fertilizer experiment was initiated in 1981 at South Lake station (30°37′N, 114°20′1″E), Hubei Province, China, where rice-wheat rotation is the common cropping system. The site is located in the northern subtropical to middle subtropical transitional geographic climate zone with an annual average total accumulated temperature of 5189.4 °C (>10 °C/day) and precipitation of 1300 mm. The tested yellow-brown paddy soil belongs to Udalfs with clay loam texture (USDA soil classification). At the beginning of the experiment in 1981, the soil had a  $pH(H_2O)$  of 6.3, organic matter of  $27.43 \,\mathrm{g \, kg^{-1}}$ , total N, P, K of  $1.801 \,\mathrm{g \, kg^{-1}}$ ,  $1.004\,g\,kg^{-1}$  and  $30.22\,g\,kg^{-1}\!,$  respectively. The concentrations of available P and K were  $5.0 \text{ mg kg}^{-1}$  and  $98.5 \text{ mg kg}^{-1}$ . Six treatments (three replicates each) were randomly implemented in 18 plots ( $40 \text{ m}^2$  each) under a rotation of winter wheat and rice. Treatments consisted of soil without fertilizer (control, CK), fertilizer N (N), fertilizer N and P (NP), fertilizer N, P and K (NPK), organic manure plus fertilizer N, P and K (NPKM) and organic manure (M). For the NPKM treatment, fertilizer N, P and K were applied in the form of urea  $(300 \text{ kg N} \text{ha}^{-1} \text{ per year})$ , superphosphate (150 kg  $P_2O_5$  ha<sup>-1</sup> per year) and potassium chloride (150 kg  $K_2$ 0 ha<sup>-1</sup> per year), respectively, while no PK or K was applied for the N and NP treatments, respectively. Organic manure was applied as pig manure (H<sub>2</sub>O 69%) with properties of 15.1 g kg<sup>-1</sup> total N, 20.8 g kg<sup>-1</sup>  $P_2O_5$  and 13.6 g kg<sup>-1</sup>  $K_2O$  (22,500 kg  $ha^{-1}$  per year).

Sixty percent of chemical fertilizers were applied to rice and the other 40% were applied during the wheat season, while organic manure was applied equally (50:50) to the two crops. All fertilizer P and K and manure during the wheat season and the rice season were applied once as basal dressing. Meanwhile 40% of fertilizer N was applied as a basal fertilizer, 40% during tillering stage and 20% during booting stage in rice season. The amounts of N fertilizer applied to wheat were 50% as basal fertilizer, 25% for overwintering period and 25% during the jointing stage. Manure and mineral fertilizers were evenly broadcasted onto the soil surface and immediately incorporated into the plowed soil (0–20 cm depth) by tillage before sowing. According to the experimental design, the nutrient application rates of the other treatments were equal to the nutrients applied in the NPKM treatment.

Soil samples from the three replicates of each treatment were collected 1 week before wheat harvesting in May 17th, 2014. Four soil cores ( $5 \times 10 \times 18$  cm) were collected at a depth of 0–20 cm from each plot. Moist soils were gently broken apart along the natural breakpoints and passed through a 5 mm sieve to remove visible organic debris. The 5 mm sieve was used rather than a 2 mm sieve because of the unique viscid characteristic of the paddy soil. If soils were forced through a 2 mm sieve, the natural structure of the soil would be destroyed. After thorough mixing, the field-moist soil was used for particle-size fractionation.

### 2.2. Fraction procedure

Soil samples were dispersed by low-energy sonication and the particle size fractions were separated by a combination of wet sieving and centrifuging as described by Stemmer et al. (1998). Briefly, the soil-water suspension was dispersed by low-energy sonication (output energy of 0.2 kJ/g) and subsequently fractionated by a combination of wet sieving and repeated centrifugation to avoid disruption of microaggregates. Finally, five fractions were obtained for each sample: large macroaggregates (>2000 µm), coarse sand-sized fraction (2000-200 µm), fine sand-sized fraction (200–63  $\mu$ m), silt-sized fraction (63–2  $\mu$ m), and clay-sized fraction (2–0.1 µm). Field-moist soils (140 g equivalent dry weight for each sample) were suspended in 400 mL of distilled water and then equally placed into four 150 mL glass beakers. The large macroaggregates, coarse and fine sand particle size fraction  $(>63 \,\mu\text{m})$  were separated by manual wet sieving with a maximum of 700 mL of cooled distilled water. Silt-sized particles were separated from the clay fraction by four centrifugation steps at  $150 \times g$  for 5 min and at 15 °C. Between each centrifugation the pellets were resuspended in water and centrifuged again to purify the silt fraction. The combined supernatants were centrifuged at  $3900 \times g$  for 30 min to obtain clay-sized particles and the resulting same size soil fractions from the glass beakers of the same sample were pooled together. The above procedures were repeated until we got enough soil samples for all the tests. The fractions were then stored at room temperature for chemical analysis, at 4°C for extracellular enzyme analysis and at -80 °C for PLFA analysis (the soil was freeze-dried before the determination of PLFAs).

### 2.3. Soil analysis

Soil organic C and total N were determined by dichromate oxidation (Kalembasa and Jenkinson, 1973) and Kjeldahl digestion (Bremner and Mulvaney, 1982), respectively.

# 2.4. Enzyme activity

The activities of all extracellular enzymes tested except phenol oxidase and peroxidase were measured using MUB-linked or AMC-

#### Table 1

Extracellular enzymes assayed in all the particle-size fractions, their enzyme commission number (EC) and corresponding substrate (L-DOPA=L-3, 4-dihydrox-yphenylalanine, 4-MUB=4-methylumbelliferyl).

Enzyme	Substrate	EC
Phosphatase	4-MUB-phosphate	3.1.3.1
Sulfatase	4-MUB-sulfate	3.1.6.1
β-Glucosidase	4-MUB-β-D-glucoside	3.2.1.21
β-Cellobiosidase	4-MUB-β-D-cellobioside	3.2.1.91
N-acetyl-glucosaminidase	4-MUB-N-acetyl-β-D-glucosaminide	3.2.1.30
β-Xylosidase	4-MUB-β-D-xyloside	3.2.1.37
α-Glucosidase	4-MUB-α-p-glucoside	3.2.1.20
Aminopeptidase	L-Leucine-7-amino-4-methylcoumarin	3.4.11.1
Phenol oxidase	l-DOPA	1.10.3.2
Peroxidase	l-DOPA	1.11.1.7

linked model substrates yielding the highly fluorescent cleavage products 4-methylumbelliferyl (MUB) or 7-amino-4-methylcoumarin (AMC) upon hydrolysis (DeForest, 2009; Saiya-Cork et al., 2002; Wittmann et al., 2004) (Table 1). The method is very sensitive and allowed a high throughput analysis of enzymatic activities (Wittmann et al., 2004). Specifically, each equivalent of 1.0 g drv mass of fresh soil was added into a 100 mL centrifuge tube. and it was homogenized with 50 mL of 50 mM acetate buffer using a polytron homogenizer, then the mixture was poured into a round wide-mouth beaker. An additional 50 mL of acetate buffer washed the centrifuge tube and was poured into the same beaker. A magnetic stirrer was used to maintain a uniform suspension. The buffer, sample suspension, 10 µM references and 200 µM substrates (Table 1) were dispensed into the wells of a black 96-well microplate according to the strict volume and order described by DeForest (2009). The microplates were covered and incubated in the dark at 25 °C for 4 h and the fluorescence quantified using a microplate fluorometer (Scientific Fluoroskan Ascent FL, Thermo) with 365 nm excitation and 450 nm emission filters (Saiya-Cork et al., 2002). The activities were expressed in units of nmol  $h^{-1}g^{-1}$ .

The non-fluorometric enzymes, phenol oxidase and peroxidase, were measured spectrophotometrically in the clear 96-well microplate using the substrate of L-3, 4-dihydroxyphenylalanine (L-DOPA). The dispensed volume and the order of buffer, sample suspension, 25 mM L-DOPA and 0.3% H<sub>2</sub>O<sub>2</sub> were the same as for the fluorometric enzymes (DeForest, 2009). The microplates were covered and incubated in the dark at 25 °C for 20 h, and the activities were assayed by measuring the absorbance at 450 nm using the microplate fluorometer and expressed in unites of  $\mu$ mol h<sup>-1</sup> g<sup>-1</sup>.

# 2.5. PLFA profiles

Differences in the microbial community and microbial biomass among the various nutrients managements were determined by phospholipid fatty acid (PLFA) analysis following the procedure described by Wu et al. (2009). Briefly, 3 g freeze-dried soil samples were used to extract the PLFAs with a single-phase mixture of chloroform:methanol:citrate buffer (15.2 mL at a 1:2:0.8 volume ratio). The extracted fatty acids in the chloroform were fractionated into neutral lipids, glycolipids, and polar lipids using a silicabonded phase column (SPE-Si, Supelco, Poole, UK) with chloroform, acetone and methanol, respectively. The recovered polar lipids were trans esterified to the fatty acid methyl esters (FAMES) by a mild alkaline methanolysis. FAMES were quantified by gas chromatograph (N6850, Agilent) and identified with an MIDI SHERLOCKS microbial identification system (Version 4.5, MIDI, Inc., Newark, DE). Nonadecanoic acid methyl ester (19:0) was added as the internal standard. Concentrations of PLFAs were expressed in units of nmol  $g^{-1}$ .

Total microbial biomass was estimated using the total concentration of PLFAs (nmol g<sup>-1</sup>). The abundance of individual PLFAs was indicated by their % mole abundance in each sample. PLFAs were divided into various taxonomic groups based on previously published PLFA biomarker data (Bossio et al., 1998; Frostegrd et al., 1993; Green and Scow, 2000). Specifically, 16:0, 17:0, 16:1ω5c, 16:1ω7c, 16:1ω9c, 17:1ω8c, 18:1ω5c, 18:1ω7c, a15:0, a17:0, cy17:0, cy19:0w8c, i14:0, i15:0, i16:0, i17:0 and i19:0 were used to represent bacterial biomarkers. The polyunsaturated PLFAs 18:2\u00fc6,9c, 18:1\u00fc9c and 18:3\u00fc6(6,9,12) (Hill et al., 2000; Dong et al., 2014; Ai et al., 2015) were chosen to indicate fungal biomarkers. The fatty acids 16:0(10Me), 17:0(10Me) and 18:0 (10Me) were considered as biomarkers of actinomycetes. We used 17:0, i14:0, i15:0, i16:0, i17:0, a15:0, a17:0 as gram-positive bacteria biomarkers; cy17:0, cy19:0w8c, 16:1w5c, 16:1w7c, 16:1w9c, 17:1ω8c, 18:1ω5c, 18:1ω7c as gram-negative bacteria biomarkers.

# 2.6. Statistical analysis

Statistical procedures (ANOVA and principal component analysis (PCA)) were carried out with SAS and Canoco for Windows (Version 4.5) softwares, respectively, and some other complemental calculations were carried out using Origin 8, Adobe Illustrator CS4 and MS Excel 2010. Redundancy analysis (RDA) was also carried by Canoco for Windows (Version 4.5) but with the Monte Carlo permutations test (499 permutations) to determine whether the microbial community composition could be correlated to enzyme activities and properties of soil. For each variable measured in the soil, the data were analyzed by one-way ANOVA using Fisher's least significant differences (LSD, P=0.05) to determine significant differences among treatment means. Two-way ANOVA was used to determine statistical differences by soil fractions and fertilizer treatments.

# 3. Results

# 3.1. Distribution and recovery of soil fractions

Physical fractionation procedures according to particle size yielded high recovery rates (95.20–97.30%, Table 2). The three main fractions of soil aggregates, obtained by ultrasonic fractionation, were  $63-2 \,\mu$ m, >2000  $\mu$ m and 2–0.1  $\mu$ m, which represented approximately 34%, 24% and 17%, respectively, of the total soil fractions. Compared with CK, the NPKM treatment had a significantly decreased proportion of the large macroaggregates (>2000  $\mu$ m) fraction and an increased proportion of the remaining soil fractions (2000–2  $\mu$ m).

# 3.2. Soil fraction C and N

Differences in SOC and total N in the different particle-size fractions were very pronounced between fertilizer treatments, with SOC concentrations in the 200-63 µm fraction nearly double that of the >200 µm fraction and triple that of the 63-0.1  $\mu$ m fraction (Fig. 1). The fine sand fraction (200–63  $\mu$ m) had a significantly higher total N concentration than the other  $(2000-200 \,\mu\text{m} > 2-0.1 \,\mu\text{m} >> 2000 \,\mu\text{m} > 63-2 \,\mu\text{m}).$ fractions NPKM and M treatments both significantly increased total N contents compared to NPK treatment in each fraction. Significant effects of fraction and fertilizer treatments (P < 0.001) on C:N ratio in particle-size fractions (Table 3) followed the expected pattern, with lower values in the smaller fractions  $(63-2 \,\mu\text{m} \text{ and } 2-0.1 \,\mu\text{m})$  compared with the larger fractions (>63 µm). Organic managements (NPKM and M) were characterized by a significant decrease in the C:N ratio in the >2000  $\mu$ m and 63–2  $\mu$ m fractions.

#### 3.3. Patterns of soil enzyme activities

The activities of ten enzymes showed different trends for the five particle-size fractions after the long-term application of inorganic and organic fertilizers (Table 4). Both soil fractions and fertilizer treatments had strong effects on soil enzyme activities, and their interaction also had a remarkable effect on the enzyme activities, with the only exception being  $\alpha$ -glucosidase, where the interaction of fraction and fertilization did not affect the activity of  $\alpha$ -glucosidase (*P*=0.0873) as assessed by two-way ANOVA (Table 3).  $\beta$ -cellobiosidase activity was most affected by particle-size fractions (*F*=790.19).

The maximum phosphatase and sulfatase activities were found in the 2–0.1 µm fraction of the organic treatments (Table 4). The  $\beta$ -glucosidase,  $\beta$ -cellobiosidase, *N*-acetyl-glucosaminidase,  $\beta$ -xylosidase,  $\alpha$ -glucosidase, aminopeptidase, phenol oxidase and peroxidase activities were significantly higher in the 200– 63 µm fraction than in the other fractions. The highest activities of these ten enzymes in each fraction were measured in the organic treatments (NPKM or M), with the exception of phosphatase, sulfatase, *N*-acetyl-glucosaminidase and  $\beta$ -xylosidase activities in the 2000–200 µm fraction which were highest in NPK treatment. Phenol oxidase and peroxidase showed minor changes in activity compared with the other hydrolytic enzymes, where the differences in activity could reach several fold among fractions.

# 3.4. Microbial communities within soil fractions (PLFA analysis)

We measured microbial composition via phospholipid fatty acid (PLFA) analysis on physically separated soil particle-size fractions. A total of 76 PLFA individuals were detected and used as measures of total microbial biomass and the relative abundance of specific microbial groups. Both particle-size fractions and fertilizer managements yielded large differences in total PLFA and different microbial groups, except for the ratio of G+:G-, which fertilization did not significantly affect (P=0.598) and the relative abundance of bacteria (P=0.1521), which were not significantly affected by the interaction of fraction and fertilization (Table 6).

Total PLFAs, which ranged from 41.70 to 205.74 nmol  $g^{-1}$ , were typically higher in the >2000  $\mu$ m and 2000–200  $\mu$ m fractions and lower in the  $63-2\,\mu\text{m}$  and  $2-0.1\,\mu\text{m}$  fractions (Table 5). At the same time, total PLFAs were significantly increased to varying extent by organic fertilization (NPKM and M), especially in the >2000  $\mu$ m fraction, where the total PLFAs were almost 4× higher than in the control and inorganic fertilization treatments (CK and N). However, the response of concentration of total PLFAs to inorganic or organic fertilizers in the 2000–200  $\mu m$  fraction was distinctly different from the responses in other fractions, where it was abundant in theNPK treatment but less abundant in the organic (NPKM and M) treatments. Relative abundance of microbial groups under six treatments in all the fractions were also calculated, which showed а tendency of

Table 2

Particle-size distribution of soils using ultrasonic fractionation under different fertilizer managements. The percentage recovery is reported for ultrasonic fractionation.

	Particle-size fractions					
Treatments	Large macroaggregates >2000 μm	Coarse sand 2000–200 µm	Fine sand 200–63 µm	Silt 63–2 µm	Clay 2-0.1 μm	Recovery
СК	25.66 ± 1.18ab	13.63 ± 0.61bc	$4.95\pm0.25b$	$33.94 \pm \mathbf{0.56bc}$	$17.12\pm0.71ab$	$95.30\pm1.85a$
N	$21.82\pm0.58b$	$11.90 \pm 0.35c$	$6.23\pm0.59 ab$	$35.89\pm0.45 ab$	$19.61 \pm 1.28a$	$96.78\pm0.41a$
NP	$26.08\pm0.44a$	$15.72\pm2.21$ ab	$5.09\pm0.18b$	$31.78 \pm 1.34c$	$16.53 \pm 1.15b$	$95.20\pm1.22a$
NPK	$27.04 \pm 0.98 a$	$12.28\pm0.43c$	$4.48\pm1.51b$	$34.48\pm0.72bc$	$16.97\pm0.51ab$	$95.25\pm0.42a$
NPKM	$17.35 \pm 1.95c$	$18.36\pm1.13a$	$7.15\pm0.33a$	$38.79 \pm \mathbf{1.05a}$	$15.65\pm0.49b$	$97.30\pm2.08a$
М	$24.05\pm0.43ab$	$18.13\pm0.84a$	$5.61\pm0.69ab$	$32.55\pm0.97c$	$15.06\pm0.65b$	$95.41\pm1.27a$

Data are means  $\pm$  standard error, n = 3. Different lower case letters indicate significant differences (P < 0.05) among fertilizations within each fraction (Fisher's LSD test).

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**Fig. 1.** Content of soil organic C, total nitrogen and C:N ratio in particle-size fractions under different treatments. Different lower case letters indicate significant differences (*P* < 0.05) among fertilizer treatments within each fraction (Fisher's LSD test).

#### Table 3

Two-way ANOVA analysis of soil organic carbon, total nitrogen and 10 extracellular enzymes activities in five soil particle-size fractions, six fertilizer treatments each with three replicates (n = 90). The data in bold indicated that enzyme was not affected by soil fractions, fertilizer treatments or their interaction (P < 0.05).

	Particle-size fraction		Fertilizer tre	atment	Particle-size frac	tion $\times$ Fertilizer treatment
	F	Р	F	Р	F	Р
Soil organic carbon	1382.46	<0.0001	83.70	<0.0001	3.89	<0.0001
Total nitrogen	480.39	< 0.0001	106.21	<0.0001	3.14	0.0003
C:N ratio	209.37	< 0.0001	6.19	<0.0001	2.10	0.0145
Phosphatase	391.48	< 0.0001	60.94	<0.0001	6.67	<0.0001
Sulfatase	272.92	< 0.0001	10.17	<0.0001	3.20	0.0003
β-Glucosidase	476.41	< 0.0001	24.85	<0.0001	2.30	0.0069
β-Cellobiosidase	790.19	< 0.0001	31.93	<0.0001	2.14	0.0126
N-acetyl-glucosamidase	504.83	< 0.0001	17.65	< 0.0001	3.51	<0.0001
β-Xylosidase	333.62	< 0.0001	21.84	< 0.0001	2.26	0.0081
α-Glucosidase	226.59	< 0.0001	19.92	< 0.0001	1.58	0.0873
Aminopeptidase	298.66	< 0.0001	127.19	<0.0001	2.58	0.0024
Phenol oxidase	64.59	< 0.0001	52.65	<0.0001	3.73	<0.0001
Peroxidase	99.57	< 0.0001	83.19	< 0.0001	4.23	<0.0001

bacteria > actinomycetes > fungi (Table 5). The relative abundance of bacteria and fungi were higher in larger fractions (>63  $\mu$ m), while actinomycetic proportion tended to be more abundant in

small fractions (<63  $\mu$ m). Small changes of relative abundance of microbial groups among treatments in the <200  $\mu$ m fraction were observed when compared to that of larger fraction (>200  $\mu$ m). The

# Table 4

Soil extracellular enzyme activities in particle-size fractions in different soils.

	Particle-size fraction								
	>2000 µm	2000–200 µm	200–63 µm	63–2 µm	2–0.1 μm				
Phosphatase									
СК	$648.72\pm65.98b$	$1313.90\pm18.57b$	$2077.85 \pm 49.53c$	$1669.34 \pm 49.00c$	$1816.69 \pm 91.27d$				
N	$780.52 \pm 81.78 b$	$1513.24 \pm 92.86ab$	$2193.54 \pm 45.29 bc$	$1830.57 \pm 49.01 bc$	$1945.65 \pm 60.58d$				
NP	$1027.09 \pm 60.43a$	$1510.05 \pm 63.64 ab$	2236.95 ± 71.71bc	1899.90 ± 73.69bc	$2837.40 \pm 96.54c$				
NPK	$1106.19 \pm 37.41a$	$1658.11 \pm 98.96a$	$2420.72 \pm 104.01 \mathrm{ab}$	2073.73 ± 53.18b	$2992.80 \pm 33.13b$				
NPKM	$1190.01 \pm 56.46a$	$1590.23 \pm 106.75a$	$2636.04 \pm 125.45a$	$2419.63 \pm 41.01a$	$3247.66 \pm 22.73a$				
М	$1163.24 \pm 44.03a$	$1587.82 \pm 49.56a$	$2619.97 \pm 220.96a$	$2405.04 \pm 141.13a$	$3148.32 \pm 75.72$ a				
	E	D	B	C	A				
Sulfatase									
CK	$13.45\pm0.75c$	$56.18\pm0.56abc$	$58.35 \pm \mathbf{3.94b}$	$30.44 \pm 2.35c$	$71.40 \pm 0.69 b$				
N	$17.53 \pm 4.09$ bc	$59.28 \pm 3.58ab$	$65.40 \pm 1.87b$	$35.28 \pm 2.14$ bc	$72.46 \pm 3.10b$				
NP	$21.27 \pm 0.96ab$	$59.78 \pm 1.35ab$	$67.11 \pm 6.20b$	$35.23 \pm 2.14bc$ $35.53 \pm 0.28bc$	$81.63 \pm 0.40$ ab				
NPK		$61.31 \pm 5.62a$	$69.04 \pm 2.74b$	$37.81 \pm 1.99$ bc					
NPKM	22.45 ± 1.23ab				$82.30 \pm 5.24$ ab				
	24.87 ± 1.53a	$47.39 \pm 0.45c$	$69.14 \pm 1.36b$	$55.90 \pm 2.73a$	$84.43 \pm 7.58ab$				
M	22.60 ± 1.75ab E	51.47 ± 2.73bc C	84.80 ± 7.21a B	50.75 ± 1.39a D	92.01 ± 7.21a A				
	2	-	2						
3-Glucosidase CK	$167.69 \pm 18.72c$	496.41 ± 13.02b	$1116.46 \pm 26.68c$	$422.68 \pm 3.27b$	219.31 ± 24.13b				
N	$218.86 \pm 61.53 \text{ bc}$	$630.65 \pm 39.88a$	$1173.25 \pm 15.69$ bc	$430.30 \pm 14.90b$	$220.83 \pm 19.67b$				
NP	$265.24 \pm 20.50$ bc	$650.05 \pm 39.88a$ $651.22 \pm 2.84a$	$1173.23 \pm 13.090$ 1189.78 $\pm 52.00$ bc	$430.30 \pm 14.900$ $448.43 \pm 11.18b$	$220.83 \pm 19.670$ 360.41 ± 21.52a				
NPK	284.78 ± 16.90ab	$679.38 \pm 25.35a$	$1203.16 \pm 112.01$ bc	$485.19 \pm 12.18b$	$367.37 \pm 19.81a$				
NPKM	369.17 ± 13.57a	632.82 ± 28.00a	$1358.94 \pm 46.67ab$	740.31 ± 33.81a	412.31 ± 8.34a				
M	375.39 ± 7.21a C	665.15 ± 21.46a B	1495.81 ± 155.35a A	777.77 ± 49.40a B	386.96 ± 19.11a C				
		d	Λ	D	C				
β-Cellobiosidase CK	$34.51 \pm \mathbf{1.24c}$	$118.95 \pm 11.28d$	391.51 ± 5.23b	$82.81\pm1.62b$	$35.24 \pm \mathbf{6.07b}$				
N	$42.57 \pm 10.81c$	$144.50 \pm 4.79$ cd	$411.25 \pm 32.90b$	$84.98 \pm 0.22b$	$37.99 \pm 5.32b$				
NP									
	$70.53 \pm 9.44b$	$154.81 \pm 2.41$ bc	$418.55 \pm 39.24b$	$92.08 \pm 1.71b$	$58.27 \pm 3.13$ ab				
NPK	72.16 ± 2.57b	$185.28 \pm 6.87 ab$	449.50 ± 21.11b	98.59 ± 4.33b	$58.47 \pm 7.25$ ab				
NPKM	$101.39 \pm 3.72a$	$193.13 \pm 19.34a$	$522.69 \pm 32.35a$	$171.34 \pm 9.70a$	73.57 ± 7.18a				
M	99.97 ± 3.38a D	208.77 ± 9.82a B	551.21 ± 25.11a	175.58 ± 4.84a C	72.08 ± 12.69a D				
	D	В	A	C	D				
N-acetyl-glucosa									
CK	$54.88 \pm 5.62 d$	$204.77\pm~37.91b$	$480.04 \pm 15.51b$	$130.82 \pm 4.58b$	$100.85\pm5.02c$				
N	$68.52 \pm 13.34 cd$	$266.15 \pm 28.25 ab$	$521.69 \pm 9.71 ab$	$132.44 \pm 4.27b$	$102.40 \pm 14.690$				
NP	$94.97 \pm 12.51c$	$266.60 \pm a17.87b$	$528.48 \pm 19.72ab$	$137.92 \pm 5.66b$	$163.44 \pm 8.16b$				
NPK	$98.74 \pm 6.19 bc$	$342.28 \pm 27.78a$	$546.50 \pm 41.46 ab$	$144.31 \pm 5.70b$	$164.68 \pm 11.96b$				
NPKM	$137.00 \pm 7.40a$	$224.37\pm19.47b$	$561.18 \pm 12.86a$	$261.47 \pm 12.66a$	198.16 $\pm$ 0.62ab				
Μ	$128.59 \pm 4.72ab$	$237.63 \pm 19.63b$	$585.91 \pm 40.06a$	$260.53 \pm 23.15a$	$220.79 \pm 24.61a$				
	D	В	А	С	D				
β-Xylosidase									
CK	$24.91 \pm 3.45c$	$98.15 \pm 7.91a$	$206.47 \pm 2.33b$	$123.15 \pm 3.19b$	$41.16 \pm 5.55b$				
N	32.83 ± 7.22bc	$102.04 \pm 5.83a$	221.60 ± 11.55b	$127.72 \pm 1.63b$	$42.87 \pm 6.08b$				
NP	$46.13 \pm 6.93b$	$102.01 \pm 3.05 \mathrm{d}$ $104.39 \pm 2.74 \mathrm{a}$	$222.40 \pm 14.54b$	$127.83 \pm 3.24b$	$62.53 \pm 3.41a$				
NPK	$46.50 \pm 1.46b$	$120.59 \pm 4.24a$	$231.86 \pm 28.57b$	$132.54 \pm 3.52b$	$65.82 \pm 4.68a$				
NPKM	67.10 ± 3.73a	$114.80 \pm 18.89a$	$281.04 \pm 6.20a$	$194.31 \pm 5.20a$	$73.94 \pm 4.79a$				
M	64.29 ± 1.97a D	118.47 ± 23.18a C	304.55 ± 25.43a A	207.23 ± 15.83a B	74.79 ± 3.20a D				
	-	-		-	-				
x-Glucosidase CK	$20.20\pm2.56c$	$65.05\pm1.93a$	117.31 ± 1.25c	$60.92\pm1.03b$	$45.38\pm4.42b$				
N	$24.54 \pm 6.81$ bc	$65.80 \pm 2.73a$	$125.01 \pm 6.26bc$	$62.82 \pm 1.43b$	$46.50 \pm 4.60b$				
NP	$32.34 \pm 0.61b$	$67.36 \pm 2.18a$	$129.51 \pm 7.62bc$	$63.90 \pm 0.47b$	$71.16 \pm 4.47a$				
NPK	$32.39 \pm 1.94b$	$74.98 \pm 5.07a$	$130.26 \pm 17.93$ bc	$66.49 \pm 0.54b$	77.37 ± 3.18a				
NPKM	45.91 ± 1.15a	$71.75 \pm 3.10a$	$146.12 \pm 2.24ab$	$96.85 \pm 4.65a$	$85.34 \pm 2.44a$				
N	43.61 ± 2.08a C	72.40 ± 3.52a B	156.71 ± 15.21a A	94.58 ± 6.98a B	82.24 ± 11.18a B				
		-		-	-				
Aminopeptidase CK		$\textbf{354.94} \pm \textbf{4.55b}$	647.96 ± 51.53b	$335.74 \pm 10.80b$	255.78 ± 16.17b				
	$170.61 \pm 12.42c$								
N	$178.10 \pm 6.96c$	$358.05 \pm 3.94b$	$683.96 \pm 45.45b$	$338.05 \pm 9.78b$	$256.19 \pm 19.641$				
NP	247.99 ± 18.49b	$373.61 \pm 19.32b$	698.75 ± 31.86b	$354.00 \pm 11.88b$	318.10 ± 17.55b				
NPK	$244.99 \pm 9.50b$	$370.91 \pm 10.90b$	$702.74 \pm 3.97b$	$363.01 \pm 13.41b$	$329.39 \pm 25.581$				
NPKM	$475.92\pm15.87a$	$526.80\pm28.08a$	$979.43 \pm 29.81a$	$672.33 \pm 39.38a$	$593.53 \pm 55.963$				
NEKIVI	437.56±35.91a	$491.81\pm39.79a$	$1036.61\pm24.34a$	$679.64 \pm 15.05a$	$519.81 \pm 52.303$				
M	D	BC	Α	В	С				
		BC	А	В	C				

#### Table 4 (Continued)

Table 5

	Particle-size fraction							
	>2000 µm	2000–200 µm	200–63 µm	63–2 μm	2–0.1 μm			
N	$1.34\pm0.07c$	$2.30\pm0.81d$	$5.27\pm0.36b$	$6.63\pm0.20d$	$5.63\pm0.66b$			
NP	$\textbf{3.25}\pm\textbf{0.65b}$	$4.52\pm0.11c$	$6.00\pm0.38b$	$7.75\pm0.26abc$	$5.94\pm0.61b$			
NPK	$\textbf{3.05}\pm\textbf{0.58b}$	$4.88\pm0.30bc$	$7.26\pm0.78b$	$7.40 \pm 0.19 bcd$	$6.24\pm0.43$ ab			
NPKM	$7.36\pm0.09a$	$6.66\pm0.72a$	$9.81\pm1.54a$	$8.18\pm0.28ab$	$7.59\pm0.54a$			
M	$6.79\pm0.12a$	$6.18\pm0.54ab$	$10.99\pm0.21a$	$8.64\pm0.56a$	$6.59\pm0.20$ ab			
	С	С	А	А	В			
Peroxidase								
СК	$4.82\pm0.30b$	$5.10\pm0.12c$	$8.71\pm0.77b$	$7.72\pm0.53c$	$2.97\pm0.31d$			
N	$5.04\pm0.53b$	$5.84 \pm 0.48 bc$	$8.77\pm0.32b$	$\textbf{7.80} \pm \textbf{0.10c}$	$3.09\pm0.26d$			
NP	$5.35\pm0.46b$	$5.91 \pm 0.24 bc$	$9.09\pm1.12b$	$9.54\pm0.43b$	5.08 ± 0.13c			
NPK	$5.68 \pm 0.32 b$	$6.39\pm0.67b$	$11.09 \pm 1.45b$	$9.61\pm0.27b$	$5.25\pm0.52c$			
NPKM	$10.86\pm0.70a$	$8.90\pm0.29a$	$14.84 \pm 1.27 a$	$11.20\pm0.57a$	$10.25\pm0.62a$			
Μ	$10.34\pm0.48a$	$8.12\pm0.33a$	$17.50\pm0.26a$	$10.61\pm0.52 ab$	$8.41\pm0.37b$			
	С	С	А	В	С			

Data are means  $\pm$  standard error, n = 3. Different capital letters indicate significant differences (P < 0.05) among the five fractions; different lower case letters indicate significant differences (P < 0.05) among six treatments within each fraction (Fisher's LSD test).

Microbial community	composition of differen	nt narticle-size fraction	s under different fertilizer	managements
- VIICTODIAL COMMUNITY	composition of anterer	1 Darticle-size traction	s under different fertilizer	managements

Microbial PLFA composition	Treatments	Particle-size fractions	5			
F		>2000 µm	2000–200 µm	200–63 µm	63–2 μm	2–0.1 μm
Total PLFAs	СК	$62.38\pm2.92c$	157.39 ± 7.15b	$103.82\pm 6.25a$	$44.64\pm 6.80b$	$47.85\pm4.06c$
$(nmol g^{-1})$	Ν	$64.19\pm4.02c$	$159.12 \pm 9.85b$	$94.81\pm5.93a$	$41.70\pm4.59b$	$50.37 \pm 2.29 \mathrm{c}$
	NP	$150.82 \pm 6.11b$	$169.37\pm4.97 ab$	$111.85\pm6.83a$	$47.19 \pm 2.48 ab$	$65.07 \pm 12.23 bc$
	NPK	$172.10 \pm 4.22b$	$188.70\pm6.10a$	$116.20 \pm 5.05a$	$45.03\pm6.99b$	$65.31 \pm 10.28 bc$
	NPKM	$205.74 \pm 9.73a$	$112.73 \pm 3.01c$	$99.58 \pm 13.07a$	$52.66\pm5.92 ab$	$94.92\pm5.77a$
	М	$202.87 \pm 9.23a$	$101.40 \pm 2.30c$	$112.34\pm2.95a$	$61.64\pm1.22a$	$90.54\pm9.06ab$
		AB	А	BC	D	CD
Bacterial PLFA	СК	$60.10\pm0.63d$	$57.93\pm0.73a$	61.66 ± 1.11a	$57.84\pm0.62a$	57.97 ± 0.70bc
(mol %)	Ν	$62.01 \pm 1.38 cd$	$59.43\pm0.43a$	$61.65 \pm 2.16a$	$55.65\pm0.44a$	$57.01 \pm 0.29c$
	NP	$64.63\pm0.17b$	$60.06 \pm 1.23a$	$64.57\pm0.28a$	$57.72\pm1.00a$	$60.91\pm0.44ab$
	NPK	$63.63 \pm 1.44 bc$	$59.06\pm0.18a$	$60.59\pm1.43a$	$56.68\pm1.39a$	$61.33\pm1.49a$
	NPKM	$67.01 \pm 0.67a$	$60.68\pm0.89a$	$61.14 \pm 1.65a$	$58.07\pm0.52a$	$60.53\pm0.57ab$
	М	$65.33\pm1.11 \mathrm{ab}$	$59.92 \pm 1.75a$	$59.99\pm2.57a$	$56.71\pm1.24a$	$59.90 \pm 1.39 abc$
		А	С	В	D	С
Fungal PLFA	СК	$2.33 \pm 0.05 c$	$8.51\pm0.12a$	$2.76 \pm 0.16 ab$	$1.99\pm0.14b$	$2.02 \pm 0.09 c$
(mol %)	Ν	$2.28\pm0.05c$	$8.88\pm0.51a$	$2.81\pm0.17ab$	$1.76\pm0.08b$	$1.96\pm0.11c$
	NP	$7.98\pm0.38b$	$9.27\pm0.33a$	$2.76\pm0.14ab$	$2.19\pm0.32ab$	$1.97\pm0.09c$
	NPK	$8.02\pm0.07b$	$9.34\pm0.14a$	$2.55\pm0.08b$	$1.96\pm0.08b$	$2.15\pm0.08 bc$
	NPKM	$8.84\pm0.27a$	$2.66\pm0.43b$	$2.95\pm0.16ab$	$2.76 \pm 0.33 a$	$2.39\pm0.17ab$
	M	$8.86\pm0.29a$	$2.27\pm0.11b$	$3.03\pm0.13a$	$2.39\pm0.13ab$	$2.68\pm0.08a$
		Α	А	В	В	В
Actinomycetic PLFA	СК	$17.55 \pm 0.24a$	$13.23\pm0.32b$	$15.88\pm0.09a$	$18.29\pm0.51a$	$18.26\pm0.94a$
	Ν	$17.74\pm0.58a$	$13.53 \pm 0.07b$	$16.07\pm0.84a$	$17.18\pm0.19ab$	$16.93\pm0.24ab$
(mol %)	NP	$15.21\pm0.67b$	$14.22\pm0.48b$	$16.91\pm0.22a$	$18.23 \pm 0.53 a$	$17.70\pm0.02ab$
	NPK	$14.19\pm0.14b$	$13.62\pm0.10b$	$15.32\pm0.27a$	$17.27\pm0.53ab$	$16.76\pm0.22b$
	NPKM	$15.71\pm0.22b$	$16.50\pm0.18a$	$16.69\pm0.72a$	$18.11\pm0.25a$	$17.57\pm0.12ab$
	Μ	$15.35\pm0.71b$	$16.41\pm0.46a$	$15.74\pm0.50a$	$16.52\pm0.45b$	$16.94\pm0.25ab$
		В	С	В	А	А

Data are means  $\pm$  standard error, n = 3. Different capital letters indicate significant differences (P < 0.05) among the five fractions; different lower case letters indicate significant differences (P < 0.05) among treatments within each fraction (Fisher's LSD test).

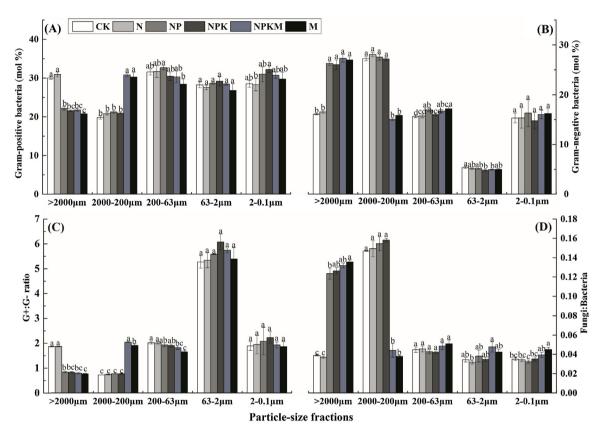
relative abundance of fungi in 2000–200  $\mu$ m fraction showed higher in inorganic treatments, while that of actinomycetes showed higher in organic treatments. The results of fungi and actinomycetes in >2000  $\mu$ m fraction were in contrary with that of 2000–200  $\mu$ m fraction. Significant increase of gram–positive bacterial relative abundance in the 2000–200  $\mu$ m fraction of organic treatments (Fig. 2A) and significant decrease of gram– negative bacterial proportion (Fig. 2B) could explain the significant increase in the G+:G– ratio in the organic treatments compared with the inorganic treatments (Fig. 2C). Additionally, relative abundance of gram-negative bacterial was excessively low in the  $63-2 \mu m$  fraction, which resulted in a remarkable increase in the G +:G- ratio in this fraction. We should note that the fungi: bacteria ratios of the >2000  $\mu m$  and 2000-200  $\mu m$  fractions showed the opposite tendency with the G+:G- ratios.

Principal component analysis (PCA) was conducted with PLFAs that were presented in the five particle-size fractions. PC1 and PC2 accounted for 49.70% and 13.10% of the total variation, respectively. PC scores on these axes were well separated on the basis of soil fractions and fertilizer treatments and the 30 plots

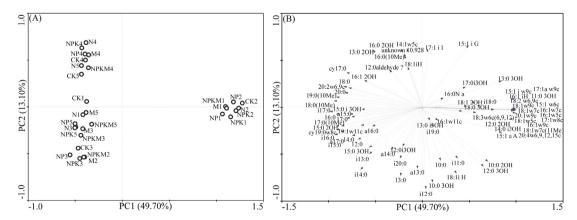
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**Table 6** Two-way ANOVA analysis of microbial groups in five soil particle-size fractions, six fertilizer treatments each with three replicates (n = 90). The data in bold indicated that microbial group was not affected by soil fractions, fertilizer treatments or their interaction (P < 0.05).

	Particle-size fraction		Fertilizer tre	atment	Particle-size fraction $\times$ Fertilizer treatment		
	F	Р	F	Р	F	Р	
Total PLFA	240.96	< 0.0001	26.83	< 0.0001	23.85	<0.0001	
Bacteria	28.17	< 0.0001	4.34	0.0020	1.41	0.1521	
Fungi	710.76	< 0.0001	39.06	< 0.0001	110.17	<0.0001	
Actinomycetes	45.92	< 0.0001	6.63	< 0.0001	5.31	<0.0001	
G+/G-	546.9	< 0.0001	0.74	0.5980	5.46	<0.0001	
Fungi/bacteria	588.87	< 0.0001	30.3	< 0.0001	92.76	<0.0001	



**Fig. 2.** Comparisons of Gram-positive bacteria (A); Gram-negative bacteria (B); G+:G- ratio (C) and Fungi: bacteria ratio (D). Vertical bars represent the S.E. (*n* = 3) and lower case letters indicate significant differences among fertilizer treatments (*P* < 0.05). G+:G- = Gram-positive bacteria:Gram-negative bacteria ratio.



**Fig. 3.** Plot of the first two principle components (PC1 and PC2) grouped in six fertilizer treatments and five particle-size fractions (A) and plot of two principle components among PLFA individuals (B) from all the fractions. Numbers (1–5) following the treatment indicate the particle sizes of >2000 µm, 200–200 µm, 200–63 µm, 63–2 µm, 2–0.1 µm, respectively.

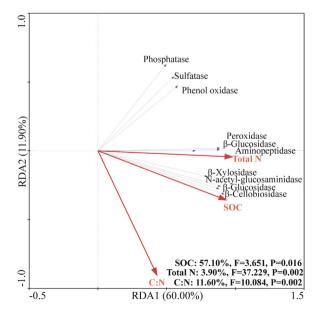
were divided into three sections (Fig. 3A). Most of the >2000  $\mu$ m and 2000–200 µm fractions were well separated along PC1. The 63–2 µm fractions and CK and N treatments of 2–0.1 µm fractions were separated as one group along PC2. All of the 200-63 µm fractions, the inorganic treatments of the >2000 µm and the other treatments of 2–0.1 µm fractions were grouped against PC1 by the principal component analysis. For PLFAs from all the fractions, the proportions of monounsaturated fatty acids  $(15:1\omega 6c, 16:1\omega 5c,$ 16:1ω7c, 16:1ω9c, 17:1ω8c, 18:1ω5c, 18:1ω7c and 18:1ω9c) and polyunsaturated fatty acids (18:3ω6c (6,9,12), 18:2ω6,9c and  $20:4\omega6,9,12,15c$ ) increased in the larger fractions (>200 µm). The proportions of saturated fatty acids (14:0, 17:0, 18:0, a15:0, a16:0, a17:0, i13:0, i14:0, i15:0, i16:0, i17:0), cyclopropane fatty acids (cy17:0 and cy19:0 $\omega$ 8c) and methyl branched fatty acids (16:0 (10Me), 17:0 (10Me), 18:0(10Me) and 19:0(10Me)) (biomarkers of actinomycetes) increased against PC1 (Fig. 3B).

#### 3.5. Correlations

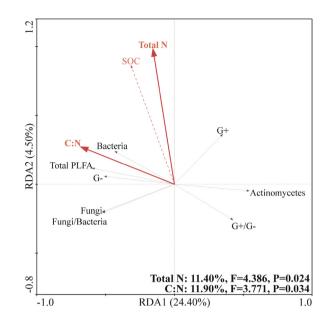
The redundancy analysis (RDA) was carried out for enzyme activities and soil properties of soil fractions of the six treatments. Soil properties were used as environmental variables (Fig. 4). The first and second axes accounted for 60.00% and 11.90% of the total variation between enzyme activities and soil C and N pools. The enzyme activities showed significant correlation with SOC (F=3.651, P=0.016), total N (F=37.229, P=0.002) contents and C:N ratio (F=10.084, P=0.002).

Results of the RDA between soil properties and microbial community composition are shown in Fig. 5. The first and second axes accounted for 24.40% and 4.50%, respectively, of the total variation in microbial community composition and structure. The microbial community composition was significantly correlated with total N content (F=4.386, P=0.024) and C:N ratio (F=3.771, P=0.034), which explained 11.40% and 11.90% of total variance, respectively. Conversely, SOC (F=1.845, P=0.196) had no significant correlation with microbial community composition of different particle-size fractions of the six fertilizer treatments.

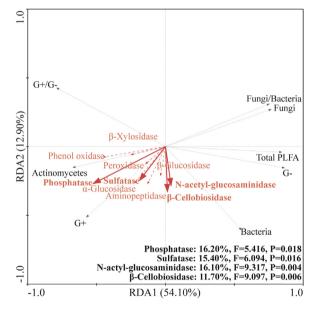
The RDA showed the relationships of soil enzyme activities with microbial community composition (Fig. 6), and the first



**Fig. 4.** Correlations of soil enzyme activities to soil properties as determined by redundancy analysis (RDA) and corresponding explained proportion of variability was shown in the lower right corner.



**Fig. 5.** Correlations of soil properties and microbial community composition indicated by PLFA groups as determined by redundancy analysis (RDA) and corresponding explained proportion of variability was shown in the lower right corner.



**Fig. 6.** Correlations of soil enzyme activities and microbial community composition indicated by PLFA groups as determined by redundancy analysis (RDA) and corresponding explained proportion of variability was shown in the lower right corner.

and second axes accounted for 54.10% and 12.90%, respectively, of the total variation. Soil enzyme activities, including phosphatase (F=5.416, P=0.018), sulfatase (F=6.094, P=0.016), N-acetyl-glucosaminidase (F=9.317, P=0.004),  $\beta$ -cellobiosidase (F=9.097, P=0.006) activities of different soil fractions were significantly correlated with microbial community after long-term fertilization. However, peroxidase (F=1.851, P=0.168),  $\beta$ -xylosidase (F=2.769, P=0.074), phenol oxidase (F=0.642, P=0.438),  $\beta$ -glucosidase (F=0.904, P=0.37), aminopeptidase (F=0.618, P=0.49) and  $\alpha$ -glucosidase (F=0.429, P=0.582) were not significantly correlated with microbial community after long-term fertilization.

# 4. Discussion

This study explored the effects of inorganic fertilizers and organic amendment on the relationships between soil microbial community composition, enzyme activities and soil C and N pools in different soil particle fractions of yellow-brown paddy soil in the Yangtze Plain region of China. We investigated soil particle-size fractions by ultrasonic fractionation. Consistent differences in biological characteristics and elemental chemistry among soil fractions and across fertilizer managements indicated that the physical fractions represented different C and N pools with potentially different turnover rates, which could lead to changes in soil biological processes. Significant benefits of organic fertilizer application were also observed. However, the drivers for the structure and activity of microbial communities appear to differ depending on the spatial scale of each experiment (Martiny et al., 2011), this is why large scale studies are required to evaluate predictors of microbial community composition and structure.

# 4.1. Fertilizer effects on soil particle-size fractions and nutrient pools

It is well known that the addition of easily decomposable substrates to soil rapidly stimulates the soil microflora and determines the proportions of different particle-size fractions, resulting in a significant improvement in soil structure (De Gryze et al., 2005; Abiven et al., 2007). In this experiment, this is confirmed by the increased proportion of the  $2000-2 \,\mu m$  fraction and the decreased proportion of the  $>2000 \,\mu m$  and  $2-0.1 \,\mu m$  fractions in the organic treatments (NPKM and M) (Table 2). Better aggregation contributes to C storage by creating a more complex soil structure and limiting C accessibility to decomposers (Smith et al., 2014).

Silveira (2005) stated SOC balance was mainly determined by the net result of carbon input and decomposition rates in soil. Our results also revealed that 33 years of organic fertilizer application (NPKM and M) to the yellow-brown paddy soil significantly increased SOC and total N contents across all the fractions. This confirmed that appropriate use of manures and/or crop residues within management systems can enhance soil microbial biomass, activity and diversity (Mandal et al., 2007), which can also increase the levels of soil nutrients in turn. At the microenvironment scale, coarse sand fractions are characterized by high concentrations of labile C and N originating predominately from plant residues, whereas silt and clay fractions are usually characterized by high concentrations of relatively stable organic C and N (Elliott, 1986; Six et al., 2000). Moreover, high SOC contents are usually found in silt and clay fractions, whereas low contents are observed in the sand fractions of most soils, such as in Calcaric Phaeozems (Kandeler et al., 2000), Calcic Chernozems and Cambisols (Stemmer et al., 1998) and Humic Dystrudepts (Chiu et al., 2006). In the present study, however, we detected the highest SOC concentrations in fine sand fraction (200–63  $\mu$ m), followed by the coarse sand fraction, which generally agrees with previous findings in Hydragric Anthrosol in which crop residues were returned to the paddy soil (Jiang et al., 2011). It has been demonstrated that the organic C (fresh or labile) derived from crop residues is first incorporated into the coarse sand fraction during the initial decomposition period and subsequently accumulates and becomes stable in silt or clay soils (Angers et al., 1997; Six et al., 2000). Consequently, it is possible that higher fresh SOC contents could first accumulate in larger fractions, particularly in soils receiving large amounts of crop residues (Chen et al., 2014). The C:N ratio indicates the degradation of fresh plant residues, which is important in the process of C sequestration (Potter et al., 1998). Organic particles of >200 µm, which consist predominantly of particulate residues from plant material, decompose very rapidly,

while the organic material in smaller size fractions is characterized by an increase in the degree of humification and a decrease in the C:N ratio (Kanazawa and Filip, 1986). It is possible that smaller fractions contain carbon and nitrogen that has undergone more degradation than the macroaggregate fraction and therefore have comparatively less diversity occluded in the microaggregate fractions contributing to the C:N ratio (An et al., 2010). This may explain the relatively high C:N ratio in the >63  $\mu$ m fraction and the low C:N ratio in the 63–0.1  $\mu$ m fraction.

# 4.2. Fertilizer effects on soil extracellular enzyme activities in soil particle-size fractions

Extracellular enzymes are unevenly distributed through the soil, and depending on their location, they may be more or less sensitive to environmental changes (Nannipieri et al., 2002). Soil particle-size fractionation is one way of dividing the bulk enzyme activity into functionally meaningful components, which is particularly useful when investigating effects of long-term management on soil enzymes (Stemmer et al., 1999). The techniques used for soil fractionation are diverse and, as a consequence, the quality and numbers of fractions generated are variable (Kandeler et al., 1999). It is therefore difficult to make comparisons between experiments. Most fractionation techniques include steps for aggregate disruption and wet sieving. As a result of these procedures, enzymes previously located in the soil solution or detached during the fractionation process may be washed out and lost from the analysis. Only enzymes that were previously immobilized (adsorbed on organic matter or clay minerals, incorporated into humic matter, or associated with living microorganisms) remain in the assay (Perez Mateos and Gonzalez Carcedo, 1985).

The general pattern of soil enzyme activities is often dominated by the amount and quality of organic substances as well as by various physical and chemical protection mechanisms (Allison and Jastrow, 2006; Nannipieri et al., 2012; Lagomarsino et al., 2012). Our findings affirmed this observation that the majority of the variation in potential enzyme activities could be explained by soil characteristics related to nutrient availability, which are well known to be strongly influenced by fertilizer managements (Table 3). In the present study, the highest enzyme activities were observed in the 200-63 µm fraction of NPKM or M treatments, except for phosphatase and sulfatase which were highest in the 2- $0.1 \,\mu m$  fraction. The 200–63  $\mu m$  sized fractions of NPKM and M soils both contained higher SOC and total N contents than the other fractions. Long-term application of organic manures expectedly contributed to increased enzyme activities with the exception of phosphatase, sulfatase and  $\beta$ -xylosidase, which showed higher activities in the 2000–200 µm fraction in the NPK treatment. This phenomenon probably occurred because those enzymes were predominantly associated with the larger fractions and therefore poorly protected during the fractionation procedure, especially under the organic treatments (Qin et al., 2010). Strong effects of soil fractions and fertilizations on activities of all tested enzymes were determined (Table 3). The results further showed that the enzyme activities of soil fractions were differentially altered by fertilization. Additionally, RDA was carried out to evaluate the correlation between soil enzyme activities and soil properties (Fig. 4). The cosine of the angle between the soil enzyme activity and soil property shows the type of relationship between these factors. SOC, total N and the ratio of C:N all showed significant positive correlations. Therefore, our results were in agreement with previous conclusions that long-term compost amendment could increase SOC by increasing organic C in all fractions, thus changing the soil enzyme activities of different soil fractions (Yu et al., 2012). Differences in enzyme activity can also depend on the type of humic compounds in soil (Benitez et al., 2005; Nannipieri et al., 2002).

# 4.3. Fertilizer effects on microbial community composition in soil particle-size fractions

Consistent trends in microbial compositional shifts and C and N pools among fertilizer treatments across different particle-size fractions further supported the strong relationships between microbial community composition and substrate microsite heterogeneity. Balanced fertilizer applications, especially with organic manure addition, enhanced the microbial biomass as reflected by total PLFA as well as PLFAs associated with specific functional groups (Table S1). This trend appeared to correspond to the significantly higher concentrations of SOC and total N under organic treatments. Total PLFAs were typically higher in the >2000  $\mu$ m and 2000–200  $\mu$ m fractions and lower in the 63–2  $\mu$ m and 2–0.1  $\mu$ m fractions. During the research of soil bonding agents, Chatterjee and Jain (1970) found microbial agents contributed to the formation of macroaggregates  $(2000-200 \,\mu\text{m})$  and the important agents of microaggregate (<200 µm) were sesquioxides of iron and aluminum. While the abundance of PLFAs in the 2000-200 µm fraction was found to be significantly higher than in the smaller fractions, the application of organic applications treatments significantly decreased the abundance of the statistical PLFA groups in the 2000–200  $\mu$ m fraction (P < 0.05) (Table S1). Previous studies found that the coarse sand fraction (2000-200 µm) was always characterized by high concentrations of labile C and N originating predominantly from plant residues and the relatively loose structure under organic treatments, which may cause active microorganisms to be easily washed away during the fractionation procedure (Qin et al., 2010; Chen et al., 2014). Additionally, bacteria dominate in rice paddy and may compete with fungi for substances, resulting in greater stress on the fungal community, which is convinced by the significant lower relative abundance of fungi under organic treatments in Table 5. The significant reduction in the fungi: bacteria ratio in the 2000–200  $\mu$ m fraction of the organic treatments compared with the NPK treatments was not surprising.

Reduced fungal abundance with decreasing soil aggregate or particle size has been documented across soil and vegetation types (Kandeler et al., 1999, 2000; Poll et al., 2003; Schutter and Dick, 2002) and has been attributed to decreased bioavailability of C substrates in the smaller size fractions (Briar et al., 2011; Chiu et al., 2006). The higher relative abundance of fungi in the larger aggregates is likely because of more favorable substrate properties (Huygens et al., 2008), higher C:N ratios, which are known to favor fungal colonization (Bossuyt et al., 2005; Eiland et al., 2001; Six et al., 2006; Waring et al., 2013), and restricted access to smaller pore sizes (Heijnen and Van Veen, 1991; Six et al., 2006). Even though the highest C:N ratio in the 200–63 µm fraction did not result in the highest fungal abundance, the relatively higher C:N ratio in the >200  $\mu$ m fraction compared with the 63–0.1  $\mu$ m fraction was in accord with the trend described above. Fungi may preferentially colonize larger soil fractions as they have been shown to grow well in soils with high porosity and low bulk density, which are facilitated by a well-developed soil structure (Harris et al., 2003). The smaller C:N ratios for the silt and claysized fractions (63–2  $\mu$ m and 2–0.1  $\mu$ m) also suggested differences in the degree of alteration of SOM and the preferential enrichment of microbial products and microbially-processed SOM in mineralassociated fractions.

Conversely, bacterial enrichment is often reported for the smallest size fractions because of the favorable bacterial microclimate (Ranjard and Richaume, 2001). We found distinct groups of bacteria dominated the different aggregate fractions. Specifically,

gram-positive bacteria were more abundant than gram-negative bacteria in silt and clay fractions, and gram-negative bacteria were more abundant in the larger fractions (>200 µm). These differences were consistent with the general understanding of grampositive and gram-negative substrate preference and observed C:N ratios of the different particle-size fractions. Gram-negative bacteria preferentially use fresh plant inputs as C sources, whereas gram-positive bacteria are thought to favor older and more microbially-processed SOM (Fierer et al., 2003; Kramer and Gleixner, 2006; Potthast et al., 2012). The greater G+:G- ratios in the silt and clay sized fractions compared with the larger fractions suggested that the silt and clay fractions are depleted in easily-decomposable substrate. Studies have shown that a higher fungi:bacteria ratio can reflect the relative abundance of the microbial population, which is an important indicator of a stronger soil ecosystem buffering capacity and more sustainable land use (Bossio et al., 1998; De Vries et al., 2006). In the present study, relatively lower G+:G- ratio and higher fungi:bacteria ratios were found in the larger fractions (>2000  $\mu$ m and 2000–200  $\mu$ m), which indicated better soil conditions and stronger ecosystem buffering capacity for crops.

PCA plots showed that the relatively more diverse microbial community in the larger fractions (>200  $\mu$ m) and the relatively less diverse microbial community in the smaller fractions  $(63-2 \mu m)$ were well separated. The microbial communities of particles of 200–63  $\mu$ m and most 2–0.1  $\mu$ m treatments together with organic treatments of 2000-200 µm were separated into one group. As mentioned previously, enzyme activities and microbial abundance were significantly decreased in the 2000–200 µm fractions when compared with other fractions and this phenomenon was confirmed by the PCA plots. As shown in Fig. 3A, the 2000-200 µm fractions of the NPKM and M treatments were grouped with the 200-63 µm and 2-0.1 µm fractions instead of being grouped with the >2000 µm and 2000-200 µm fractions of the other treatments along PC1. PCA analysis also indicated fatty acids, which mostly were biomarkers of gram-negative bacteria and fungi, were positively correlated with PC1and strongly affected >2000  $\mu$ m and 2000–200  $\mu$ m fractions. This result was convinced by the higher relative abundance of fungi and gram-negative bacteria of organic treatments of >2000 µm and inorganic treatments of 2000-200 µm fraction. Fatty acids, which mostly represented gram-positive bacteria and actinomycetes, were negatively correlated with PC1. The treatments, affected by gram-positive bacteria and actinomycetes in PCA plot (Fig. 3A), contained higher proportion of these microbial groups (Table 5 and Fig 2A).

RDA was also carried out to detect the relationships between soil properties, enzyme activities and microbial community composition (Figs. 5 and 6). Total N, C:N ratio, phosphatase, sulfatase, N-acetyl-glucosaminidase and B-cellobiosidase activities were significantly correlated with microbial community composition after long-term fertilization, whereas no significant correlations were detected between SOC, phenol oxidase,  $\beta$ -xylosidase, peroxidase,  $\beta$ -glucosidase, aminopeptidase and  $\alpha$ -glucosidase activities with microbial community composition. Ling et al. (2014) stated that these enzyme activities might be affected by specific functional microbial communities rather than by the overall microbial community. It is worth noting that phosphatase, sulfatase and N-acetyl-glucosaminidase activities, which were significantly related to the microbial community composition, were lower in the organic treatments but higher in the NPK treatment of 2000–200 µm fraction. This may partly explain the decline in microbial biomass in the 2000–200 µm fraction of the organic treatments. The consistent changes again confirmed the strong relationship between enzyme activities and microbial communities.

The high variability in results among studies may be because of soil-specific properties such as mineralogy and differences in soil C and nutrient content (Chiu et al., 2006), or differences in the methods used to characterize microbial biomass and community composition. Most studies on the relationship between soil physical fractions and microbial communities have been conducted in temperate, agricultural soils, so their applicability to other ecosystems is limited. Further work across environmental gradients and a diversity of soil types will increase understanding of the relationship between soil structure and microbial biomass, composition and, ultimately, function.

# 5. Conclusion

Soil particle-size fractions and fertilizer regimes individually and interactively (P < 0.05) impacted soil carbon and nitrogen contents, enzyme activities and microbial communities of yellowbrown paddy soil in China. The long-term application of organic fertilizer contributed to the improvement of soil C, N contents and most enzyme activities. The highest SOC, total N and activities of most enzymes were existed in 200-63 µm fraction. Larger fractions (>63 µm) held higher microbial abundance and better soil conditions for microorganisms than smaller sizes as reflected by phospholipid fatty acid analysis. Strong relationships between total N, C:N ratio, phosphatase, sulfatase, N-acetyl-glucosaminidase and  $\beta$ -cellobiosidase activities with soil microbial community composition were also found. Thus, we recommend soil fractionation as a promising approach that offers potential in analyzing the relationship between soil functional diversity and soil fertility. Nevertheless, further information is necessary about functions of the observed uncultured bacteria and the relationship between functional microbial activity and enzyme activities.

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#### Appendix A. Supplementary data

Supplementary data associated with this article can be found, in the online version, at http://dx.doi.org/10.1016/j. apsoil.2015.05.005.

#### References

- Abiven, S., Menasseri, S., Angers, D.A., Leterme, P., 2007. Dynamics of aggregate stability and biological binding agents during decomposition of organic materials. Eur. J. Soil Sci. 58, 239–247.
- Ai, C., Liang, G., Sun, J., He, P., Tang, S., Yang, S., Zhou, W., Wang, X., 2015. The alleviation of acid soil stress in rice by inorganic or organic ameliorants is associated with changes in soil enzyme activity and microbial community composition. Biol. Fertil. Soils 1–13.
- Allison, S.D., Jastrow, J.D., 2006. Activities of extracellular enzymes in physically isolated fractions of restored grassland soils. Soil Biol. Biochem. 38, 3245–3256.
- An, S., Mentler, A., Mayer, H., Blum, W.E., 2010. Soil aggregation, aggregate stability, organic carbon and nitrogen in different soil aggregate fractions under forest and shrub vegetation on the Loess Plateau, China. CATENA 81, 226–233.
- Angers, D.A., Recous, S., Aita, C., 1997. Fate of carbon and nitrogen in water-stable aggregates during decomposition of <sup>13</sup>C<sup>15</sup>N-labelled wheat straw in situ. Eur. J. Soil. Sci. 48, 295–300.
- Benitez, E., Sainz, H., Nogales, R., 2005. Hydrolytic enzyme activities of extracted humic substances during the vermicomposting of a lignocellulosic olive waste. Bioresour. Technol. 96, 785–790.
- Bernard, E., Larkin, R.P., Tavantzis, S., Erich, M.S., Alyokhin, A., Sewell, G., Lannan, A., Gross, S.D., 2012. Compost rapeseed rotation, and biocontrol agents

significantly impact soil microbial communities in organic and conventional potato production systems. Appl. Soil Ecol. 52, 29–41.

- Bossio, D.A., Scow, K.M., Gunapala, N., Graham, K.J., 1998. Determinants of soil microbial communities: effects of agricultural management, season, and soil type on phospholipid fatty acid profiles. Microb. Ecol. 36, 1–12.
- Bossuyt, H., Six, J., Hendrix, P.F., 2005. Protection of soil carbon by microaggregates within earthworm casts. Soil Biol. Biochem. 37, 251–258.
- Bremner, J.M., Mulvaney, C.S., 1982. Nitrogen-total. In: Page, A.L., Miller, R.H., Keeney, D.R. (Eds.), Methods of Soil Analysis: Chemical and Microbiological Properties Part 2. ASA, Madison, pp. 643–698.
- Briar, S.S., Fonte, S.J., Park, I., Six, J., Scow, K., Ferris, H., 2011. The distribution of nematodes and soil microbial communities across soil aggregate fractions and farm management systems. Soil Biol. Biochem. 43, 905–914.
- Burns, R.G., 1982. Enzyme activity in soil: location and a possible role in microbial ecology. Soil Biol. Biochem. 14, 423–427.
- Cavigelli, M.A., Mirsky, S.B., Teasdale, J.R., Spargo, J.T., Doran, J., 2013. Organic grain cropping systems to enhance ecosystem services. Renew. Agric. Food Syst. 28, 145–159.
- Chatterjee, R.K., Jain, J.K., 1970. Studies on aggregate formation with reference to cementing substances. Soil Sci. Plant Nutr. 16, 231–233.
- Chen, J., He, F., Zhang, X., Sun, X., Zheng, J., Zheng, J., 2014. Heavy metal pollution decreases microbial abundance: diversity and activity within particle-size fractions of a paddy soil. FEMS Microbiol. Ecol. 87, 164–181.
- Chiu, C.Y., Chen, T.H., Imberger, K., Tian, G., 2006. Particle size fractionation of fungal and bacterial biomass in subalpine grassland and forest soils. Geoderma 130, 265–271.
- Das, B., Chakraborty, D., Singh, V.K., Aggarwal, P., Singh, R., Dwivedi, B.S., Mishra, R.P., 2014. Effect of integrated nutrient management practice on soil aggregate properties: its stability and aggregate-associated carbon content in an intensive rice-wheat system. Soil Till. Res. 136, 9–18.
- De Gryze, S., Six, J., Brits, C., Merckx, R., 2005. A quantification of short-term macroaggregate dynamics: influences of wheat residue input and texture. Soil Biol. Biochem. 37, 55–66.
- De Vries, F.T., Bardgett, R.D., 2012. Plant-microbial linkages and ecosystem nitrogen retention: lessons for sustainable agriculture. Front. Ecol. Environ. 10, 425–432.
- De Vries, F.T., Hoffland, E., Van Eekeren, N., Brussaard, L., Bloem, J., 2006. Fungal/ bacterial ratios in grasslands with contrasting nitrogen management. Soil Biol. Biochem. 38, 2092–2103.
- De Vries, F.T., Manning, P., Tallowin, J.R.B., Mortimer, S.R., Pilgrim, E.S., Harrison, K.A., Hobbs, P.J., Quirk, H., Shipley, B., Cornelissen, J.H.C., Kattge, J., Bardgett, R.D., 2012. Abiotic drivers and plant traits explain landscape-scale patterns in soil microbial communities. Ecol. Lett. 15, 1230–1239.
- DeForest, J., 2009. The influence of time storage temperature, and substrate age on potential soil enzyme activity in acidic forest soils using MUB-linked substrates and L-DOPA. Soil Biol. Biochem. 41, 1180–1186.
- Degens, B.P., Harris, J.A., 1997. Development of a physiological approach to measuring the catabolic diversity of soil microbial communities. Soil Biol. Biochem. 29, 1309–1320.
- Degens, B.P., Schipper, L.A., Sparling, G.P., Vojvodic-Vukovic, M., 2000. Decreases in organic C reserves in soils can reduce the catabolic diversity of soil microbial communities. Soil Biol. Biochem. 32, 189–196.
- Denef, K., Six, J., 2005. Clay mineralogy determines the importance of biological versus abiotic processes for macroaggregate formation and stabilization. Eur. J. Soil Sci. 56, 469–479.
- Denef, K., Six, J., Bossuyt, H., Frey, S.D., Elliott, E.T., Merckx, R., Paustian, K., 2001. Influence of dry–wet cycles on the interrelationship between aggregate particulate organic matter, and microbial community dynamics. Soil Biol. Biochem. 33, 1599–1611.
- Dong, W.Y., Zhang, X.Y., Dai, X.Q., Fu, X.L., Yang, F.T., Liu, X.Y., Sun, X.M., Wen, X.F., Schaeffer, S., 2014. Changes in soil microbial community composition in response to fertilization of paddy soils in subtropical China. Appl. Soil Ecol. 84, 140–147.
- Eiland, F., Klamer, M., Lind, A.M., Leth, M., Baath, E., 2001. Influence of initial C/N ratio on chemical and microbial composition during long term composting of straw. Microb. Ecol. 41, 272–280.
- Elliott, E.T., 1986. Aggregate structure and carbon nitrogen, and phosphorus in native and cultivated soils. Soil Sci. Soc. Am. J. 50, 627–633.
- Fierer, N., Schimel, J.P., Holden, P.A., 2003. Variations in microbial community composition through two soil depth profiles. Soil Biol. Biochem. 35, 167–176.
- Fierer, N., Strickland, M.S., Liptzin, D., Bradford, M.A., Cleveland, C.C., 2009. Global patterns in belowground communities. Ecol. Lett. 12, 1238–1249.
- Frostegrd, A., Bååth, E., Tunlio, A., 1993. Shifts in the structure of soil microbial communities in limed forests as revealed by phospholipid fatty acid analysis. Soil Biol. Biochem. 25, 723–730.
- Gattinger, A., Muller, A., Haeni, M., Skinner, C., Fliessbach, A., Buchmann, N., Mäder, P., Stolze, M., Smith, P., Scialabba, N.E.-H., Niggli, U., 2012. Enhanced top soil carbon stocks under organic farming. Proc. Natl. Acad. Sci. U. S. A. 109, 18226–18231.
- Green, C., Scow, K., 2000. Analysis of phospholipid fatty acids (PLFA) to characterize microbial communities in aquifers. Hydrogeol. J. 8, 126–141.
- Harris, K., Young, I.M., Gilligan, C.A., Otten, W., Ritz, K., 2003. Effect of bulk density on the spatial organisation of the fungus Rhizoctonia solani in soil. FEMS Microbiol. Ecol. 44, 45–56.
- Heijnen, C.E., Van Veen, J.A., 1991. A determination of protective microhabitats for bacteria introduced into soil. FEMS Microbiol. Lett. 85, 73–80.

- Hill, G.T., Mitkowski, N.A., Aldrich-Wolfe, L., Emele, L.R., Jurkonie, D.D., Ficke, A., Nelson, E.B., 2000. Methods for assessing the composition and diversity of soil microbial communities. Appl. Soil. Ecol. 15, 25–36.
- Huygens, D., Denef, K., Vandeweyer, R., Godoy, R., Cleemput, O., Boeckx, P., 2008. Do nitrogen isotope patterns reflect microbial colonization of soil organic matter fractions? Biol. Fertil. Soils 44, 955–964.
- Jackson, L.E., Bowles, T.M., Hodson, A.K., Lazcano, C., 2012. Soil microbial-root and microbial-rhizosphere processes to increase nitrogen availability and retention in agroecosystems. Curr. Opin. Environ. Sustain. 4, 517–522.
- Jiang, X., Wright, A.L., Wang, X., Liang, F., 2011. Tillage-induced changes in fungal and bacterial biomass associated with soil aggregates: a long-term field study in a subtropical rice soil in China. Appl. Soil Ecol. 48, 168–173.
- Kalembasa, S.J., Jenkinson, D.S., 1973. A comparative study of titrimetric and gravimetric methods for the determination of organic carbon in soil. J. Sci. Agric. 24, 1085–1090.
- Kanazawa, S., Filip, Z., 1986. Distribution of microorganisms total biomass, and enzyme activities in different particles of brown soil. Microb. Ecol. 12, 205–215.
- Kandeler, E., Palli, S., Stemmer, M., Gerzabek, M.H., 1999. Tillage changes microbial biomass and enzyme activities in particle-size fractions of a Haplic Chernozem. Soil Biol. Biochem. 31, 1253–1264.
- Kandeler, E., Tscherko, E.D., Bruce, K.D., Stemmer, M., Hobbs, P.J., Bardgett, R.D., Amelung, W., 2000. Structure and function of the soil microbial community in microhabitats of a heavy metal polluted soil. Biol. Fertil. Soils 32, 390–400. Kennedy, A.C., Gewin, V.L., 1997. Soil microbial diversity: present and future
- considerations. Soil Sci. 162, 607–617. Kramer, C., Gleixner, G., 2006. Variable use of plant-and soil-derived carbon by
- microorganisms in agricultural soils. Soil Biol. Biochem. 38, 3267–3278.
- Lagomarsino, A., Grego, S., Kandeler, E., 2012. Soil organic carbon distribution drives microbial activity and functional diversity in particle and aggregate-size fractions. Pedobiologia 55, 101–110.
- Ling, N., Sun, Y., Ma, J., Guo, J., Zhu, P., Peng, C., Yu, G., Ran, W., Guo, S., Shen, Q., 2014. Response of the bacterial diversity and soil enzyme activity in particle-size fractions of Mollisol after different fertilization in a long-term experiment. Biol. Fertil. Soils 1–11.
- Liu, E., Yan, C., Mei, X., He, W., Bing, S.H., Ding, L., Liu, Q., Liu, S., Fan, T., 2010. Longterm effect of chemical fertilizer, straw, and manure on soil chemical and biological properties in northwest China. Geoderma 158, 173–180.
- Mandal, A., Patra, A.K., Singh, D., Swarup, A., Masto, R.E., 2007. Effect of long-term application of manure and fertilizer on biological and biochemical activities in soil during crop development stages. Bioresour. Technol. 98, 3585–3592.
- Martiny, J.B.H., Eisen, J.A., Penn, K., Allison, S.D., Horner-Devine, M.C., 2011. Drivers of bacterial b-diversity depend on spatial scale. Proc. Natl. Acad. Sci. 108, 7850– 7854.
- Mikha, M.M., Rice, C.W., 2004. Tillage and manure effects on soil and aggregateassociated carbon and nitrogen. Soil Sci. Soc. Am. J. 68, 809–816.
- Moeskops, B., Buchan, D., Sleutel, S., Herawaty, L., Husen, E., Saraswati, R., Setyorini, D., De Neve, S., 2010. Soil microbial communities and activities under intensive organic and conventional vegetable farming in West Java, Indonesia. Appl. Soil Ecol. 45, 112–120.
- Moeskops, B., Buchan, D., Van Beneden, S., Fievez, V., Sleutel, S., Gasper, M.S., D'Hose, T., De Neve, S., 2012. The impact of exogenous organic matter on SOM contents and microbial soil quality. Pedobiologia 55, 175–184.
- Nannipieri, P., Giagnoni, L., Lagomarsino, G., Puglisi, E., Ceccanti, B., Masciandaro, G., Fornasier, F., Moscatelli, M.C., Marinari, S., 2012. Soil enzymology: classical and molecular approaches. Biol. Fertil. Soils 48, 743–762.
- Nannipieri, P., Kandeler, E., Ruggiero, P., 2002. Enzyme activities and microbiological and biochemical processes in soil. Enzymes in the Environment. Marcel Dekker, New York, pp. 1–33.
- Neufeldt, H., Ayarza, M.A., Resck, D.V., Zech, W., 1999. Distribution of water-stable aggregates and aggregating agents in Cerrado Oxisols. Geoderma 93, 85–99. Nsabimana, D., Haynes, R.J., Wallis, F.M., 2004. Size, activity and catabolic diversity
- of the soil microbial biomass as affected by land use. Appl. Soil Ecol. 26, 81–92. Pankhurst, N.W., Purser, G.J., Van Der Kraak, G., Thomas, P.M., Forteath, G.N.R., 1996.
- Effect of holding temperature on ovulation, egg fertility, plasma levels of reproductive hormones and in vitro ovarian steroidogenesis in the rainbow trout *Oncorhynchus mykiss*. Aquaculture 146, 277–290.

- Perez Mateos, M., Gonzalez Carcedo, S., 1985. Effect of fractionation on location of enzyme activities in soil structural units. Biol. Fertil. Soils 1, 153–159.
- Poll, C., Theide, A., Wermbter, N., Sessitsch, A., Kandeler, E., 2003. Micro-scale distribution of microorganisms and microbial enzyme activities in a soil with long term organic amendment. Eur. J. Soil Sci. 54, 715–724.
- Potter, K.N., Torbert, H.A., Jones, O.R., Matocha, J.E., Morrison Jr., J.E., Unger, P.W., 1998. Distribution and amount of soil organic C in long-term management systems in Texas. Soil Till. Res. 47, 309–321.
- Potthast, K., Hamer, U., Makeschin, F., 2012. Land-use change in a tropical mountain rainforest region of southern Ecuador affects soil microorganisms and nutrient cycling. Biogeochemistry 111, 151–167.
- Qin, S., Hu, C., He, X., Dong, W., Cui, J., Wang, Y., 2010. Soil organic carbon: nutrients and relevant enzyme activities in particle-size fractions under conservational versus traditional agricultural management. Appl. Soil. Ecol. 45, 152–159.
- Ranjard, L., Richaume, A., 2001. Quantitative and qualitative microscale distribution of bacteria in soil. Res. Microbiol. 152, 707–716.
- Saiya-Cork, K.R., Sinsabaugh, R.L., Zak, D.R., 2002. The effects of long term nitrogen deposition on extracellular enzyme activity in an Acer saccharum forest soil. Soil Biol. Biochem. 34, 1309–1315.
- Schutter, M.E., Dick, R.P., 2002. Microbial community profiles and activities among aggregates of winter fallow and cover-cropped soil. Soil Sci. Soc. Am. J. 66, 142– 153.
- Seufert, V., Ramankutty, N., Foley, J.A., 2012. Comparing the yields of organic and conventional agriculture. Nature 485, 229–232.
- Silveira, M.L.A., 2005. Dissolved organic carbon: bioavailability of N and P as indicators of soil quality. Sci. Agric. 62, 502–508.
- Six, J., Frey, S.D., Thiet, R.K., Batten, K.M., 2006. Bacterial and fungal contributions to carbon sequestration in agroecosystems. Soil Sci. Soc. Am. J. 70, 555–569.
- Six, J., Elliott, E.T., Paustian, K., 2000. Soil macroaggregate turnover and microaggregate formation: a mechanism for C sequestration under no-tillage agriculture. Soil Biol. Biochem. 32, 2099–2103.
- Smith, A.P., Marín-Spiotta, E., de Graaff, M.A., Balser, T.C., 2014. Microbial community structure varies across soil organic matter aggregate pools during tropical land cover change. Soil Biol. Biochem. 77, 292–303.
- Stemmer, M., Gerzabek, M.H., Kandeler, E., 1998. Organic matter and enzyme activity in particle-size fractions of soils obtained after low-energy sonication. Soil Biol. Biochem. 30, 9–17.
- Stemmer, M., Gerzabek, M.H., Kandeler, E., 1999. Invertase and xylanase activity of bulk soil and particle-size fractions during maize straw decomposition. Soil Biol. Biochem. 31, 9–18.
- Tang, J., Mo, Y., Zhang, J., Zhang, R., 2011. Influence of biological aggregating agents associated with microbial population on soil aggregate stability. Appl. Soil. Ecol. 47, 153–159.
- Tisdall, J.M., 1994. Possible role of soil microorganisms in aggregation in soils. Plant Soil 159, 115–121.
- Wallenstein, M.D., Burns, R.G., 2011. Ecology of extracellular enzyme activities and organic matter degradation in soil: a complex community-driven process. Methods Soil Enzymol, 35–55.
- Wardle, D.A., Ghani, A., 1995. A critique of the microbial metabolic quotient (*q* CO<sub>2</sub>) as a bioindicator of disturbance and ecosystem development. Soil Biol. Biochem. 27, 1601–1610.
- Waring, B.G., Averill, C., Hawkes, C.V., 2013. Differences in fungal and bacterial physiology alter soil carbon and nitrogen cycling: insights from meta-analysis and theoretical models. Ecol. Lett. 16, 887–894.
- Wittmann, C., Kähkönen, M., Ilvesniemi, H., Kurola, J., Salkinoja-Salonen, M., 2004. Areal activities and stratification of hydrolytic enzymes involved in the biochemical cycles of carbon nitrogen, sulphur and phosphorus in podsolized boreal forest soils. Soil Biol. Biochem. 36, 425–433.
- Wu, Y., Ding, N., Wang, G., Xu, J., Wu, J., Brookes, P.C., 2009. Effects of different soil weights, storage times and extraction methods on soil phospholipid fatty acid analyses. Geoderma 150, 171–178.
- Yu, H.Y., Ding, W.X., Luo, J.F., Donnison, A., Zhang, J.B., 2012. Long-term effect of compost and inorganic fertilizer on activities of carbon-cycle enzymes in aggregates of an intensively cultivated sandy loam. Soil Use Manage. 28, 347–360.